

Comparing the ZnO/Fe(VI), UV/ZnO and UV/Fe(VI) processes for removal of Reactive Blue 203 from aqueous solution

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Abstract

Background: Wastewater contaminated with dyes such as Reactive Blue 203 can produce a lot of health problems if it is released into the environment without a suitable treatment. Although there are several studies on dye removal from wastewater, removal of Reactive Blue 203 has not been investigated by hybrid methods. Therefore, the aim of this study was to investigate the removal of Reactive Blue 203 from aqueous solution, using combined processes of zinc oxide (ZnO) nanoparticles, Fe(VI) oxidation process, and UV radiation.

Methods: The removal of dye from aqueous solution using ZnO nanoparticles, Fe(VI) oxidation process, and UV radiation was individually evaluated. Then, the results of combined methods were compared. Hydraulic retention time (HRT), pH, and temperature were the most important factors which were investigated in this study.

Results: ZnO nanoparticles, Fe(VI) oxidation process, and UV radiation were able to remove 97%, 71%, and 47% of the dye in the optimal conditions, respectively. Also, the removal of dye using combination of Fe(VI) oxidation process/UV radiation, ZnO nanoparticles/Fe(VI) oxidation process, and ZnO nanoparticles/UV radiation under optimum conditions was 100%. It seems that the combined methods were significantly more effective than the methods alone for removal of dye from water.

Conclusion: UV radiation alone is a simple and efficient method for removal of Reactive Blue 203 from water. Removal of Reactive Blue 203 using Fe(VI) oxidation process can be completed in a fraction of second, therefore, it can be categorized as a rapid reaction.

Keywords: Wastewater, Ultraviolet rays, Zinc oxide, Adsorption

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Introduction

Widespread use of chemicals in various industries in the world results in the production of a large amount of wastewater (1). There is a large amount of chemicals in industrial wastewater. These chemical compounds can be toxic, carcinogenic, teratogenic, and mutagenic (2). Therefore, it is necessary to treat wastewater before its release in the environment (3). Dyes are chemical compounds that make the wastewater colored. The use of dyes for textile fabrics turns back to 10 000 BC. The Reactive Blue 203 dye is one of the most common used dyes in the textile industry and its molecular structure is

shown in Figure 1.

Although dyes are usually produced in the form of solid powder, they should be dissolved into the aquatic solution for painting the fabric. Therefore, textile industries are responsible for producing a large amount of contaminated wastewater that can increase serious environmental issues (4). Most dyes, even in extreme light conditions, are very stable. For this reason, they cannot be effectively eliminated by physical, chemical, and biological methods (5). A wide range of dye concentration between 30 and 714 mg/L has been reported in different studies (6,7). Different methods such as oxidation, adsorption, and



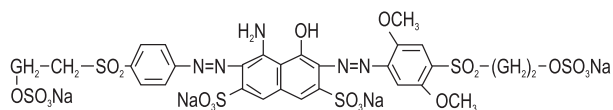


Figure 1. Molecular structure of Reactive Blue 203.

photo-oxidation processes are usually used to remove dye from wastewater. Oxidation processes are considered as a popular method for the treatment of water-soluble dyes (8). Ozone, chlorine, chlorine dioxide, and hydrogen peroxide are among the well-known oxidants used to treat colored wastewater. Although chlorine is an inexpensive oxidizer, trihalomethanes (THMs) are formed as a byproduct in the wastewater treatment (9). Since THMs have carcinogenic effects, they are considered as the environmental pollutants. Therefore, the use of chlorine as an oxidizing agent can remove a wide range of pollutants and generate new pollutants in wastewater. Ozone and chlorine dioxide have no side effects but they are very expensive (10). Recent studies have introduced new oxidizing agents that can be used in the wastewater treatment. Fe(VI) is a chemical that has the highest oxidation power in the acidic environment compared to the other oxidants (11). This chemical not only oxidizes organic compounds, but also kills microorganisms (12). In aqueous solutions, Fe(VI) converts to Fe(III), which is known as a coagulant (10). Fe(VI) can be used for oxidation, coagulation, and disinfection of urban and industrial wastewater in one reactor. Many attempts have been made to use the Fe(VI) oxidation process to remove various contaminants from the wastewater (11,13-21).

Photocatalytic oxidation is another popular method that has been widely used to remove dye from wastewater. Photocatalytic oxidation is defined as chemical change through light radiation, including visible and invisible wavelengths (22). Nowadays, researchers are studying nanosemiconductor photocatalysts since they have several exciting properties (23). In many cases, particles between 1 to 100 nm in size are called nanoscale particles. Solid particles are composed of atoms and molecules. By reducing the particle size, solid particles exhibit different characteristics, probably due to changes in the bonding state of the atoms and molecules of the particles. The nanomaterials have a large surface area and are totally different from bulk materials (24). Therefore, the nanomaterials show incredible efficiency for different uses (25). There are many known semiconductor such as zinc oxide (ZnO) and titanium oxide (TiO₂), which have a large energy gap. Such semiconductors show appropriate results for different applications like paints, gas sensor, biosensor, sunscreen, and solar cells (26,27). From commercial aspect, ZnO has priority to other semiconductors as it is very cheap. For this reason, it was used in this study. Nanoparticles of ZnO are prepared by several methods including alkaline deposition, thermal decomposition, micro-emulsion, organic-zinc hydrolysis,

and thermal degradation of the plasma. Some studies claimed that existing ZnO nanoparticles can enhance some certain removal processes of pollutants from water (22,28). Darvishi Cheshmeh Soltani et al reported that the treatment of textile wastewater can be improved by using ZnO nanoparticles (29). Darvishi Cheshmeh Soltani et al used a rectangular photo-reactor equipped with carbon black (CB)/ZnO nanocomposite film to degrade methyl orange dye (30). They found that the optimum initial dye concentration, reaction time, CB/ZnO ratio and initial pH were 13 mg/L, 95 minutes, 0.05, and 5, respectively. They reported that under optimum condition nearly 80% of methyl orange dye could be removed by CB/ZnO. The ability of ZnO nanoparticles can be enhanced by ultraviolet radiation. UV is a powerful radiation that is applied for wastewater treatment alone or with photocatalytic oxidants (31). Talaiekhosani et al reported that UV radiation can be efficiently used to remove hydrogen sulfide from municipal wastewater (32). Using UV radiation with photocatalytic oxidants such as ZnO or TiO₂ can improve pollutants removal (33). UV causes a large number of chemical interactions. These interactions are more intense at a wavelength about 0.3 μm.

Although many studies have already been done on the removal of pollutants using UV radiation, Fe(VI) oxidation process, and ZnO nanoparticles (34-39), but no study has investigated the effect of combination of these methods to remove dye from wastewater. In this study, the effects of Fe(VI) oxidation process, ZnO nanoparticles, and UV radiation and their combination on the removal of Reactive-Blue 203 dye were investigated.

Materials and Methods

Analytical methods

A synthetic wastewater was used for the experiments. To prepare synthetic wastewater, a suitable amount of Reactive Blue 203 dye was dissolved into distilled water to provide a solution of 39 mg/L. All of the chemicals except Reactive Blue 203 and ZnO nanoparticles were purchased from Merck Company. Reactive Blue 203 was purchased from Alwan Company (Hamedan, Iran). ZnO nanoparticles were purchased from Scharlau Company (Barcelona, Spain) with a purity of 99.9%. Several studies have shown that Reactive Blue 203 can be measured by a spectrophotometer (40). Therefore, the concentration of dye in synthetic wastewater was measured using a UV-VIS spectrophotometer (UNICO model S2100 SUV) at a wavelength of 631 nm. To measure the amount of UV radiation, a GBT5/OFRG3 8W UV meter was used. In this study, for all experiments, equation 1 was used to determine the removal efficiency of dye.

$$RE = \frac{C_1 - C_2}{C_1} \times 100 \quad (1)$$

where C_1 is the initial concentration of dye (mg/L), C_2 is the secondary concentration of dye (mg/L), and RE is the

removal efficiency of dye from synthetic wastewater (%). The experiments were performed as batch and one factor at a time (OFAT) method. Scanning electron microscope (SEM), X-ray diffraction (XRD), and Brunauer-Emmett-Teller (BET) test are usually used to analyze some characteristics such as size and specific surface area of nanoparticles (5). In this study, the characteristics of ZnO nanoparticles were analyzed using SEM, X-RD, and BET.

UV radiation

Since UV radiation cannot pass through the glass, a dish ($20 \times 12 \times 2.5$ cm) was used to study the dye removal using UV radiation, which is called UV container in this study. The UV container had a low depth and high surface. The UV lamp was installed in a distance of few millimeters of the synthetic wastewater surface.

pH

Seven UV containers containing 100 mL of synthetic wastewater were prepared. Then, pH of synthetic wastewater in the containers was adjusted to 1, 3, 5, 7, 9, 11, and 13 by adding a suitable amount of 4 N hydrochloric acid or 2 N sodium hydroxide. Next, the containers were placed under UV radiation with power of 170 mW/cm^2 at 23°C for 15 minutes. At the end, the dye concentration in the containers was measured.

Hydraulic retention time

Fourteen UV containers containing 100 mL of synthetic wastewater were prepared. Then, the pH of synthetic wastewater was adjusted to 13 by adding a suitable amount of 2 N sodium hydroxide. It should be noted that in this study, the optimum pH for dye removal using UV radiation was obtained to be 13, therefore, this pH was used for the next experiments. Next, the containers were exposed to UV radiation with power of 170 mW/cm^2 at 23°C for 1 to 55 minutes. Eventually, the dye concentration in the containers was determined.

Temperature

Six UV containers containing 100 mL of synthetic wastewater were prepared. Then, pH of synthetic wastewater was adjusted to 13. Next, the containers were exposed to UV radiation with power of 170 mW/cm^2 for 50 minutes at temperatures of 10, 30, 40, 50, 55, and 60°C , respectively. Finally, the concentration of remaining dye in the containers was measured.

UV radiation

Five UV containers containing 100 mL of synthetic wastewater were prepared. Then, pH was adjusted to 13 by 2 N sodium hydroxide. Next, the containers were exposed to UV radiation with power of 10 to 170 mW/cm^2 at 23°C for 15 minutes. After that, the dye concentration in each container was measured.

Fe(VI) oxidation process

There are several effective factors on Fe(VI) oxidation process such as concentration of organic matter, pH, Fe(VI) concentration, temperature, presence of impurities etc (12,16-18). In this study, the effect of pH, Fe(VI) concentration, hydraulic retention time (HRT), and temperature was investigated.

pH

Eleven 250 mL Erlenmeyer flasks containing 100 mL of synthetic wastewater were prepared. Then, pH of synthetic wastewater was adjusted to 1 to 13 by adding a suitable amount of hydrochloric acid or sodium hydroxide, respectively. Next, Fe(VI) was added to each flask to obtain the concentration of 3.3 mg/L. After that, the flasks were kept at 23°C for 15 min. Finally, the concentration of remaining dye in the solutions was determined.

Fe(VI) concentration

Six 250 mL Erlenmeyer flasks containing 250 mL of synthetic wastewater were prepared. Then, pH of synthetic wastewater was adjusted to 1 by adding a suitable amount of hydrochloric acid. Next, 0.9, 1.5, 2, 2.7, 3.3, and 4 mg of Fe(VI) was added to each flask, respectively. After that, the flasks were kept at 23°C for 15 min. Eventually, the dye concentration in the flasks was measured.

Hydraulic retention time

Eleven 250 mL Erlenmeyer flasks containing 100 mL of synthetic wastewater were prepared. Then, pH of synthetic wastewater was adjusted to 1 by adding a suitable amount of hydrochloric acid. Next, Fe(VI) was added to the flasks to obtain the concentration of 3.3 mg/L. After that, the flasks were kept at 23°C for 1 to 40 minutes, respectively. Finally, the concentration of remaining dye in each flask was determined.

Temperature

Six 250 mL Erlenmeyer flasks containing 100 mL of synthetic wastewater were prepared. Then pH of the synthetic wastewater in each flask was adjusted to 1 by adding a suitable amount of hydrochloric acid. Next, a suitable amount of Fe(VI) was added to each flask to obtain the concentration of 3.3 mg/L. After that, the flasks were kept at temperatures of 23, 30, 40, 50, 60, and 70°C for 15 minutes. Finally, the dye concentration in each flask was measured.

ZnO nanoparticles

pH

Six 250 mL Erlenmeyer flasks containing 100 mL of synthetic wastewater were prepared. Then, pH of each flask was adjusted to 1, 3, 6, 9, 11, and 13, respectively. Next, ZnO nanoparticles were added to each flask to achieve the concentration of 200 mg/L. Next, the flasks were incubated by a shaker incubator at 150 rpm at 23°C

for 15 minutes. Finally, the dye concentration in each flask was measured.

ZnO concentration

Nine 250 mL Erlenmeyer flasks containing 100 mL of synthetic wastewater were prepared. Then, a suitable amount of ZnO nanoparticles was added to each flask to obtain the concentration of 200 to 1000 mg/L, respectively. Next, pH was adjusted to 9 by adding sodium hydroxide. After that, the flasks were incubated by a shaker incubator at 150 rpm at 23°C for 15 minutes. Ultimately, the concentration of remaining dye in each flask was determined.

Hydraulic retention time

Six 250 mL Erlenmeyer flasks containing 100 mL of synthetic wastewater were prepared. Then, pH of synthetic wastewater was adjusted to 9 by adding a suitable amount of sodium hydroxide. After that, ZnO nanoparticles were added to each flask to achieve the concentration of 800 mg/L. After that, the flasks were incubated by a shaker incubator at 150 rpm at 23°C for 1 to 30 minutes, respectively. Then, the concentration of remaining dye in each flask was measured.

Temperature

Seven 250 mL Erlenmeyer flasks containing 100 mL of synthetic wastewater were prepared. Then, Fe(VI) was added to each flask to obtain the concentration of 800 mg/L. Next, the pH was adjusted to 9 by adding a suitable amount of sodium hydroxide. After that, the flasks were incubated by a shaker incubator at 150 rpm for 15 minutes at temperatures of 10, 20, 30, 40, 50, 60, and 70°C, respectively. Finally, the concentration of remaining dye in the flasks was measured.

Isotherm models

In this study, isotherm models of Langmuir, Freundlich, Temkin, Dubinin–Radushkevich (D-R), Generalized, and Jovanovic were investigated (Table 1). Freundlich isotherm has been empirically achieved. Freundlich isotherm is shown in Eq. 2. In this equation, x/m is the amount of adsorbed adsorbate per adsorbent, C_e is the equilibrium concentration of adsorbate in the solution after adsorption process, and n and K_f are constant coefficients of Freundlich isotherm. Eq. (2) is analogous of linear equation ($y = ax + b$). The amount of K_f and n is calculated using plotting $\log(x/m)$ versus $\log(C_e)$. Langmuir isotherm is shown in Eq. 3. In this equation, a and b are constant coefficients of Langmuir isotherm. Eq. 3 is similar to the general linear equation; therefore, a and b can be calculated by plotting $C_e(x/m)$ versus C_e . The Temkin isotherm is shown in Eq. 4. In this equation, K_T is the equilibrium binding constant (L/mg) and constant B_1 is related to heat of adsorption. In this equation, the constants K_T and B_1 can be calculated using a linear plot of q_e versus $\ln(C_e)$. The D-R isotherm is shown in Eq. 5. In this equation, q_e is D-R constant and ε can be calculated using Eq. 6. In Eq. 5 and 6, q_e is the maximum amount of adsorbate that can be adsorbed on the adsorbent, B is the constant related to energy, C_e is the equilibrium concentration (mg/L), R is the universal gas constant that is equal to 8.314 J/mol.K and T is temperature (Kelvin). The Generalized isotherm is shown in Eq. 7. In this equation, K_G is the saturation constant (mg/L), N is the cooperative binding constant, q_{max} is the maximum adsorption capacity of the adsorbent (mg/g), q_e (mg/g) and C_e (mg/L) are the equilibrium dye concentrations in the soil and liquid phase, respectively. The values of N and K_G in Eq. 7 are obtained from the slope and intercept of the plots. Eq. 8 shows the linear form of the Jovanovic model, where

Table 1. The list of isotherms investigated in this study

Isotherm Equations	Isotherm	Equation No.
$\log\left(\frac{x}{m}\right) = \log K_f + \frac{1}{n} \log(C_e)$	Freundlich	(2)
$\frac{C_e}{(x/m)} = \frac{1}{ab} + \frac{1}{a} C_e$	Langmuir	(3)
$q_e = B_1 \ln(K_T) + B_1 \ln(C_e)$	Temkin	(4)
$\ln q_e = \ln q_s - B \varepsilon^2$	D-R (Part 1)	(5)
$\varepsilon = RT \ln\left(1 + \frac{1}{C_e}\right)$	D-R (Part 2)	(6)
$\ln\left[\left(\frac{q_{max}}{q_e}\right) - 1\right] = \ln(K_G) - N \ln(C_e)$	Generalized	(7)
$\ln q_e = \ln q_{max} - K_J C_e$	Jovanovic	(8)

C_e is the equilibrium concentration (mg/L), K_f is constant coefficient of Jovanovic, q_e is the amount of adsorbate that was adsorbed onto the adsorbent at equilibrium stage (mg/g), and q_{max} is the maximum adsorption capacity obtained from the plot of $\ln q_e$ versus C_e .

Results

UV radiation

The effect of HRT, UV radiation power, pH, and temperature on the dye removal using UV radiation is shown in Figure 2. The results showed that the increase of pH had a negative effect on the dye removal efficiency (Figure 2a). The increased temperature between 10 and 60°C did not have a significant effect on the dye removal (Figure 2b). The dye removal efficiency was sharply increased when HRT increased between 1 and 15 minutes (Figure 2c). The UV power between 10 and 70 mW/cm² had a significant effect on the dye removal (Figure 2d).

Fe(VI)

The effect of different factors including pH, HRT, temperature, and concentration of Fe(VI) on the dye removal using Fe(VI) oxidation process is shown in Figure 3. The results showed that the dye removal efficiency using Fe(VI) sharply decreased when pH increased from 1 to 4 and 10 to 13 (Figure 3a). But the dye removal efficiency had a constant trend at pH values between 4 and 10. The reaction between Fe(VI) and the dye could be completed

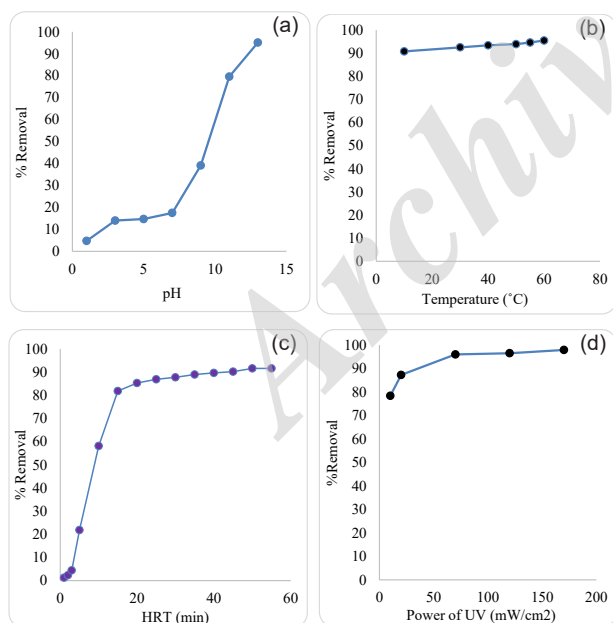


Figure 2. The effect of HRT, UV radiation power, pH, and temperature on dye removal using UV radiation. (a) The effect of pH on dye removal using UV radiation at 23°C, UV radiation of 170 mW/cm², and HRT of 15 min. (b) The effect of temperature on dye removal efficiency using UV radiation at pH 13, UV radiation of 170 mW/cm², and HRT of 50 min. (c) The effect of HRT on dye removal using UV radiation at 23°C, pH 13, and UV radiation of 170 mW/cm². (d) The effect of UV radiation power on dye removal efficiency using UV radiation at 50°C, pH 13, and HRT of 50 min.

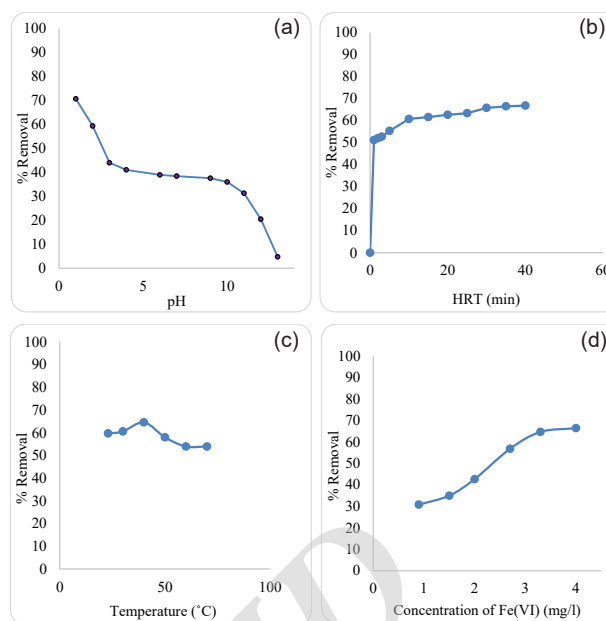


Figure 3. The effect of different factors on dye removal using Fe(VI) oxidation process. (a) The effect of pH on dye removal using Fe(VI) at 23°C, HRT of 15 min, and Fe(VI) concentration of 3.3 mg/L. (b) The effect of HRT on dye removal using Fe(VI) at 23°C, pH of 1, and Fe(VI) concentration of 3.3 mg/L. (c) The effect of temperature on dye removal using Fe(VI) at pH of 1, HRT of 10 min, and Fe(VI) concentration of 3.3 mg/L. (d) The effect of Fe(VI) concentration on dye removal using Fe(VI) oxidation process at 40°C, HRT of 10 min, and pH of 1.

within the first few seconds, therefore, HRT was not considered as an effective factor in the dye removal by Fe(VI) (Figure 3b). Although some researchers reported that the reaction between Fe(VI) and some special dye is rapid (2), but others reported that pollutants removal using Fe(VI) needs a long HRT to be completed (10). Therefore, further studies on the effect of HRT on various pollutants oxidation by Fe(VI) are required. Temperatures between 23°C and 30°C were not effective in the dye removal using Fe(VI). The maximum dye removal of 65% was obtained at 40°C. Temperature more than 40°C reduced the dye removal (Figure 3c). The increase of Fe(VI) concentration up to 3.3 mg/L had a positive effect on the dye removal (Figure 3d). Similar results about the effect of temperature on pollutants removal using Fe(VI) have been reported (2,32,39,41).

ZnO nanoparticles

In this study, SEM was used to determine the size of nanoparticles. The results showed that the size of ZnO nanoparticles was between 27 and 58 nm (Figure 4). Also, the results of X-RD test for ZnO nanoparticles are presented in Figure 5. The results of BET test revealed that the specific surface area of purchased ZnO nanoparticles was about 13 m²/g. The effect of temperature, HRT, pH, and concentration of ZnO nanoparticles on the dye removal using ZnO nanoparticles is shown in Figure 6. According to this figure, all mentioned factors had significant effect on the dye removal efficiency.

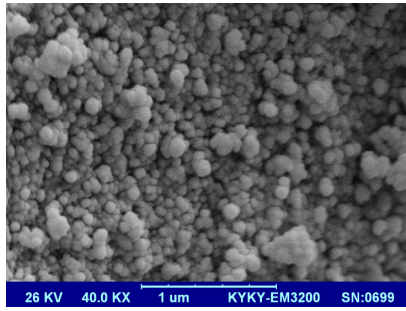


Figure 4. The SEM image of ZnO nanoparticles.

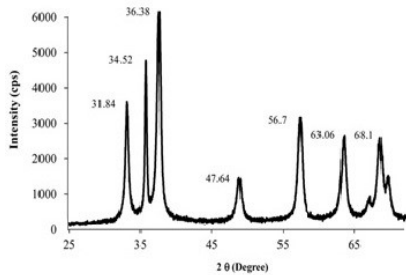


Figure 5. The results of X-RD for ZnO nanoparticles.

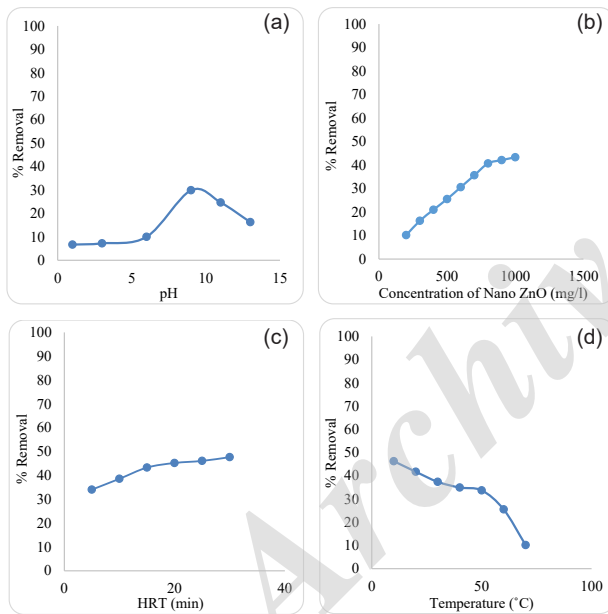


Figure 6. The effect of temperature, HRT, pH, and concentration of ZnO nanoparticles on dye removal using ZnO nanoparticles. (a) The effect of pH on dye removal efficiency using ZnO nanoparticles at 23°C, HRT of 15 min, and ZnO nanoparticles concentration of 200 mg/L. (b) The effect of ZnO nanoparticles on dye removal efficiency at 23°C, HRT of 15 min, and pH 9. (c) The effect of HRT on dye removal efficiency using ZnO nanoparticles at 23°C, pH 9, and ZnO nanoparticles concentration of 800 mg/L. (d) The effect of temperature on dye removal efficiency using ZnO nanoparticles at pH 9, HRT of 15 min, and ZnO nanoparticles concentration of 800 mg/L.

Isotherm models

Isotherm models are mathematical equations that express the equilibrium values of a substance at different concentrations that is chemically or physically adsorbed on the surface of certain solid at a constant temperature.

The investigation of different isotherm models (Figure 7) showed that the absorption of dye on ZnO nanoparticles could be described by Langmuir isotherm ($R^2=0.89$). The constant coefficients of these models are also shown in Table 2. As shown in this table, the maximum adsorption capacity calculated using Jovanovic, D-R, and Langmuir is 14.05, 22.29, and 26.59 mg/g, respectively. Since the results of this study showed that the adsorption of dye can be described by Langmuir isotherm, the real maximum capacity of ZnO nanoparticles is 26.59 mg/g (Table 2). Honorio et al reported that the maximum adsorption capacity of 57.473 mg/g could be obtained for Reactive Blue dye when soybean hulls was used as adsorbent (42). This report revealed that many other cheap and accessible adsorbents can be used instead of ZnO nanoparticles with the same and even more removal efficiency.

Combination of ZnO nanoparticles and UV radiation

One of the aims of this study was to evaluate the combination of different methods to remove dye from water. In this section of study, combination of ZnO nanoparticles with UV radiation was investigated. The

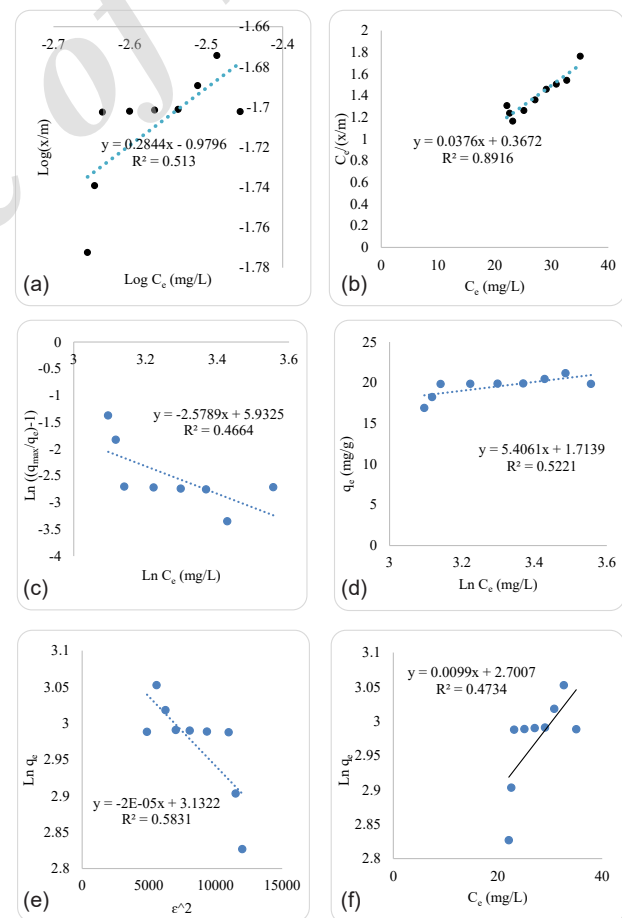


Figure 7. The regression results of Langmuir, Freundlich, Generalized, D-R, Temkin, and Jovanovic isotherm models. (a) Freundlich. (b) Langmuir. (c) Generalized. (d) Temkin. (e) D-R. (f) Jovanovic.

Table 2. Constant coefficients of Langmuir, Freundlich, Generalized, D-R, Temkin, and Jovanovic isotherm models

Jovanovic		Generalized		D-R		Temkin		Freundlich		Langmuir							
q_{max}	k_j	R^2	N	k_G	R^2	q_e	B	R^2	K_f	n	R^2	a (q_{max} (mg/g))	b (K_f (L/mg))	R^2			
14.05	0.0099	0.47	2.57	377.09	0.46	22.92	2E-05	0.58	1.37	5.40	0.52	0.10	3.51	0.51	26.59	0.102	0.89

results are shown in Figures 8 and 9. Since, the effect of nanoparticles concentration was investigated in several different studies (43), therefore, it was not considered in this study.

Combination of UV radiation and Fe(VI)

The effect of temperature, pH, and HRT on dye removal efficiency using combination of UV radiation and Fe(VI) oxidation process is illustrated in Figure 8. The results demonstrated the dye removal efficiency using this combined method at pH values between 2 and 13 was higher than pH 9 (Figure 10a). In this section of study, the effect of temperature (30 to 60°C) on the combination of UV radiation and Fe(VI) was investigated. Increasing temperature had a somewhat negative effect on the dye

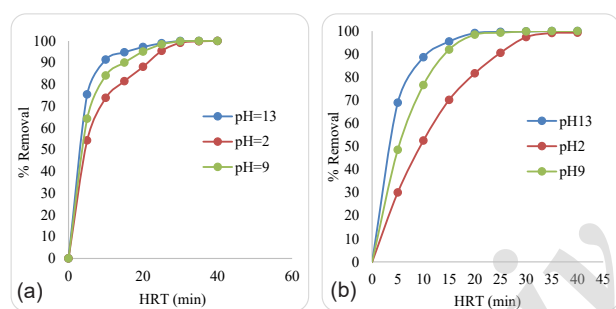


Figure 8. The effects of pH, HRT, and UV power on dye removal efficiency using combination of ZnO nanoparticles and UV radiation. (a) The effects of pH and HRT on dye removal efficiency using combination of ZnO nanoparticles and UV radiation at 23°C, UV power of 70 mW/cm², and ZnO nanoparticles concentration of 800 mg/L. (b) The effects of pH and HRT on dye removal efficiency using combination of ZnO nanoparticles and UV radiation at 23°C, UV power of 170 mW/cm², and ZnO nanoparticles concentration of 800 mg/L.

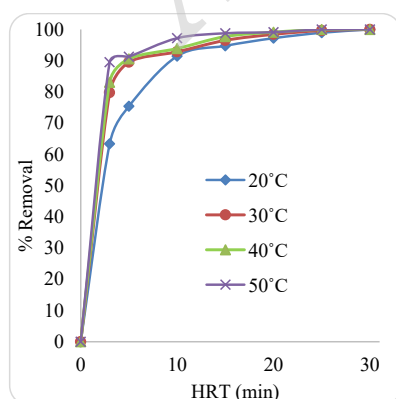


Figure 9. Effect of HRT and temperature on dye removal using combination of ZnO nanoparticles and UV radiation (70 mW/cm²).

removal using combined methods (Figure 10b).

Combination of ZnO nanoparticles and Fe(VI) oxidation process

The effect of pH, temperature, and HRT on dye removal using combination of ZnO nanoparticles and Fe(VI) oxidation process was investigated in this part of the study. The effect of pH between 2 and 13 is shown in Figure 11a. The results of evaluation of the effect of temperature on the dye removal efficiency using combination of ZnO nanoparticles and the Fe(VI) oxidation process are shown in Figure 11b.

Discussion

UV radiation

UV radiation is a powerful beam that can decompose various chemical compounds. In this study, direct UV was radiated to the surface of water contaminated with dye. The results showed that four factors of temperature, pH, HRT, and UV radiation power were effective in removing dye from contaminated water exposed to the UV radiation. pH is a factor affecting the efficiency of chemical and photochemical reactions. This study showed that increased pH values could increase the dye removal efficiency by UV radiation (Figure 2a). At pH 1, only 5% of the dye was removed from the water. By increasing pH to 7, the dye removal efficiency reached 17%. By changing the conditions from neutral to alkaline, the dye removal efficiency was increased sharply. At pH values of 9, 11, and 13, the dye removal efficiency was 39%, 80%, and

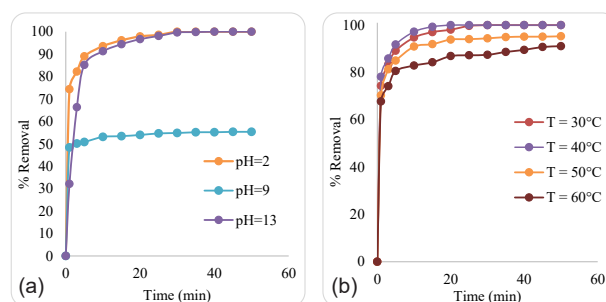


Figure 10. Effect of temperature, pH, and HRT on dye removal efficiency using combination of UV radiation and Fe(VI) oxidation process. (a) The effect of HRT on dye removal efficiency using combination of UV radiation and Fe(VI) oxidation process at different pHs, 23°C, Fe(VI) concentration of 3.3 mg/L, and UV radiation power of 170 mW/cm². (b) The effect of HRT on dye removal efficiency using combination of UV radiation and Fe(VI) oxidation process at various temperatures, pH 2, Fe(VI) concentration of 3.3 mg/L, and UV radiation power of 170 mW/cm².

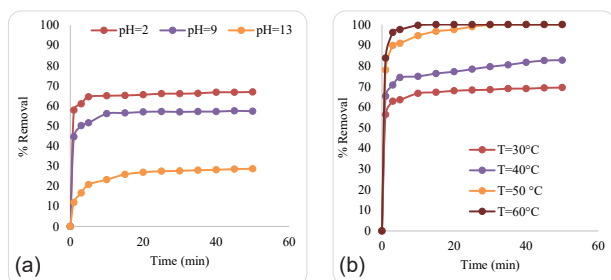


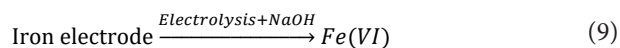
Figure 11. The effect of temperature, pH, and HRT on dye removal efficiency using combination of ZnO nanoparticles and Fe(VI) oxidation process. (a) The effect of pH and HRT on dye removal efficiency using combination of ZnO nanoparticles and Fe(VI) oxidation process at 23°C, Fe(VI) concentration of 3.3 mg/L, and ZnO nanoparticles concentration of 800 mg/L. (b) The effect of temperature and HRT on dye removal efficiency using combination of ZnO nanoparticles and Fe(VI) oxidation process at pH 2, Fe(VI) concentration of 3.3 mg/L, and ZnO nanoparticles concentration of 800 mg/L.

95%, respectively. According to Figure 2a, UV can remove the dye due to the production of free electron or free radical on the dye molecule. So, a negative charge will be produced on the dye molecules. In acidic ambient (pH<4), the breakdown of dye molecule is not possible due to the absorption of UV photons by the high concentration of H_3O^+ cations. As shown in Figure 2a, the dye removal by UV radiation was enhanced by increased pH of solution, increased OH^- concentration, and decreased H_3O^+ disturbance.

Figure 2b illustrates that the efficiency of UV to remove dye was increased at higher temperatures. So, the increase of entropy of system and the number of collision to achieve equilibrium is the main reason to enhance the UV efficiency. It was revealed that 87% of the dye was removed in the first 20 minutes of the reaction in the presence of UV (Figure 2c). As mentioned above and shown in Figure 2d, the UV radiation can remove dye with a high efficiency. The UV has been used to remove different pollutants. Talaiekhazani et al reported that nearly 100% of formaldehyde was removed from water using UV radiation (38). Such reports show that the UV radiation can be efficiently used for the removal of a wide variety of chemicals.

Fe(VI)

pH is one of the most important parameters affecting oxidation reactions. The effect of pH variations on the dye removal efficiency is shown in Figure 3a. As can be seen, by increasing pH from 1 to 13, the dye removal efficiency decreased sharply from 71% to 5%. The dye removal efficiency was constant at pH values between 3 and 9. By increasing pH from 9 to 13, the dye removal efficiency was significantly reduced. It implies that in acidic environments, the rate of removal process is higher. This fact can be explained by examining the rate of ferrate(VI) degradation. Fe(VI) is generated using electrolysis process based on equation 9.



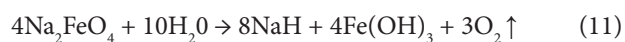
The color of electrolyte is changed to purple when Fe(VI) is generated. There are two important forms of Fe(VI): FeO_4^{2-} and $HFeO_4^-$. The rate of Fe(VI) consumption can be calculated using equation 10:

$$K(\text{Fe(VI)}) = K_1(HFeO_4^-) + K_2(FeO_4^{2-}) \quad (10)$$

Comparison of constant consumption rates of $HFeO_4^-$ ($K_1=1.24 \times 10^7$ M/S) and FeO_4^{2-} ($K_2=8.41 \times 10^2$ M/S) shows that $HFeO_4^-$ is dominant. Consequently, the reaction rate of $HFeO_4^-$ and dye was higher than that of FeO_4^{2-} (16). Therefore, $HFeO_4^-$ has a key role in the removal of dye compared with FeO_4^{2-} . Under acidic condition, the dominant species of Fe(VI) is $HFeO_4^-$ and under basic conditions, Fe(VI) is mainly appeared in the form of FeO_4^{2-} . Since $HFeO_4^-$ is more effective than FeO_4^{2-} , Fe(VI) is much more effective in acidic condition.

HRT is an important parameter in determining the size of chemical reactors. The effect of HRT changes on dye removal is shown in Figure 3b. As can be seen, the reaction between Fe(VI) and the dye was rapid. A rapid reaction means that it can be performed in a very short HRT. One of the advantages of rapid reaction is that to remove dye, a very small reactor is needed that reduces costs. Talaiekhazani et al reported that changes in HRT did not have any significant effect on Fe(VI) removal efficiency (32).

Temperature is a parameter that influences the oxidation reactions. Figure 3c shows the effect of temperature on the dye removal efficiency. As shown in this figure, the increase of temperature from 23°C to 40°C has a positive effect on the dye removal efficiency using Fe(VI). The results of this study showed that temperatures above 40°C had a negative effect on the removal of dye from the water using Fe(VI), which is consistent with the results of the study by Eskandari (10). She investigated the effect of temperature on the conversion rate of Fe(VI) to Fe(III). She reported that Fe(VI) is gradually converted to Fe(III) based on the equation 11 (10):



Eskandari demonstrated that the conversion rate of Fe(VI) to Fe(III) at temperatures between 20 and 50°C does not have a significant variation. When temperature increases to more than 50°C, the conversion rate of Fe(VI) to Fe(III) also raises, therefore, the concentration of Fe(VI) in the solution declines. Fe(III) is a strong coagulant but it cannot be used as an oxidant for the removal of organic compounds (9). It means that the ability of Fe(VI) oxidation is decreased at temperatures above 50°C. There are two known ways for degradation of the Fe(VI): (a) self-degradation and (b) reactions with organic

compounds. Talaiekhosani et al provided an empirical equation to predict the removal efficiency of dye in various environmental conditions including temperature, pH, and HRT when Fe(VI) is used (2). Although this empirical equation is specifically designed to remove 1,9-Dimethyl-Methylene Blue zinc chloride double salt, it can also be used to predict the suppression of other dyes by Fe(VI). One of the important parameters in controlling the dye removal is concentration of Fe(VI). In this study, the effect of initial concentration of Fe(VI) on the dye removal efficiency was investigated (Figure 3d). The results showed that by increasing the concentration of Fe(VI) from 0.9 to 3.3 mg/L, the dye removal efficiency increased. However, at a concentration more than 3.3 mg/L, no change in the efficiency was observed. In this study, it was found that the best ratio of dye to Fe(VI) is 11.81. Han et al reported that 97.5% bisphenol A (BPA) removal was achieved when Fe(VI) oxidation process was used (44). Such studies illustrate that Fe(VI) oxidation process can be applied for a wide range of chemicals specially for recalcitrant compounds. Zhou and Jiang revealed that Fe(VI) had a much higher reactivity with ciprofloxacin than ibuprofen (45), therefore, it can be concluded that the reaction of Fe(VI) with all chemicals is not the same. The results of this study showed that Fe(VI) oxidation is not an appropriate process for Reactive Blue 203 dye since only 71% of the dye is removed under optimum condition.

ZnO nanoparticles

Figure 4a shows that increasing the concentration of nanoparticles between 0.02 and 800 mg/L, leads to a linear increase in the removal efficiency of the dye. However, the rate of dye removal decreased when the concentration of nanoparticles increased to more than 800 mg/L. The results showed that using 200 mg/L of ZnO nanoparticle, only 10% of the dye was removed. Also, 41% of the dye was removed when the concentration of ZnO nanoparticles was 800 mg/L. Increasing the concentration of nanoparticles from 800 to 100 mg/L, resulted in only a 2% increase in the dye removal efficiency.

The amount of dye removal was increased by changing pH from acidic to alkaline. The dye removal was reduced at pH 11. It means that the effect of the ZnO nanoparticles decreased at a pH higher than 10 (Figure 4b). It can be due to this fact that the dye has a negative charge in alkaline solution and is repulsed by negative adsorber surface at pH values higher than 9. It is obvious that the concentration of OH⁻ will be saturated at a pH near to 14 and causes a disturbance on the surface of adsorber to prevent the adsorption of dye by ZnO. So, the dye removal decrease at pHs higher than 11. As shown in Figures 5c and 5d, with an increase in temperature, the efficiency of dye removal was reduced which is against the mechanism of the effect of increase of temperature and increase of entropy on adsorption. The reason of these phenomena

can be explained according to the effect of an increase in temperature on the structure of ZnO adsorber. The volume of pores and total volume of ZnO decrease by an increase in temperature due to empty space of ionized oxygen in the ZnO structure and electron transfer among stimulated pores. By increasing the volume, it can be expected that the surface of adsorption and the adsorption efficiency are decreased (46-49).

Combination of ZnO nanoparticles and UV radiation

One of the aims of this study was to evaluate the combination of different methods to remove dye from water. In this section of the study, combination of ZnO nanoparticles with UV radiation was investigated. The results showed that HRT had a major effect on the removal of dye when a combination of ZnO nanoparticles and UV radiation was used. The results showed that at a pH more than 9, the UV radiation power of 170 mW/cm², and HRT of 20 minutes approximately 98% of the dye was removed. Increasing HRT to more than 20 minutes did not increase the dye removal efficiency. The results demonstrated that pH has a key role in this process. As shown in Figure 8a, an increase in pH from 2 to 13 leads to an increase in the dye removal efficiency. At pH 2, an HRT more than 30 minutes was required to remove 98% of the dye. The effect of intensity of UV radiation on the removal efficiency of dye was also investigated in this study. The amount of UV radiation in Figures 8a and b was 170 and 70 mW/cm², respectively. The required HRT to remove 98% of the dye at pH 9 under UV radiation with power of 70 mW/cm², was 25 minutes. (i.e., 5 minutes more than that required for UV radiation of 170 mW/cm²). Further studies are needed to understand whether increasing UV radiation power is economic or increasing HRT.

In this study, temperature was also investigated and it was found that the removal efficiency of dye using combination of ZnO nanoparticles and UV radiation had a higher performance at higher temperatures. However, at temperatures above 20°C, the dye removal efficiency was very close to each other (Figure 9). The results of this study showed that at 50°C, only 98% of the dye can be removed in the HRT of 10 minutes. Further studies are required to understand increasing temperature is economic to remove dye using combination of ZnO nanoparticles and UV radiation or increasing HRT. The results showed that under UV radiation, the reaction of ZnO nanoparticles to temperature was changed. The increase in temperature has a detrimental effect on the dye removal efficiency when ZnO nanoparticles alone were used. Also, temperature was not recognized as an effective factor when UV radiation alone was used. The results revealed that when ZnO nanoparticles used under UV radiation, the temperature became an effective factor on dye removal.

The ZnO nanoparticles that draw the researchers'

attentions as a good semiconductor in the presence of UV radiation to remove pollutants from wastewater, act as a catalyst and adsorb high energized photon. So, the hydroxyl group will be produced by the nanoparticles in the presence of UV radiation. When nanoparticles adsorb photon containing energy higher than the energy gap of nanoparticle structure, it can lead to transfer of electron from the valence band to conduction band. If the separation of electrical charge happens, electron will transfer to the surface of catalyst and takes part in the reduction-oxidation reactions.

Temperature acts as an important parameter that can affect the photocatalytic adsorption. The increase of temperature makes the transfer of electrons to the surface of catalyst to be higher and the production of pore-electron to be easier. These electrons and pores themselves, are able to degrade the organic pollutants and increase the efficiency of photocatalytic pollutants removal (50). Shahrezaei et al stated that temperature was the most effective parameter on the efficiency of photocatalytic pollutants removal (50). They also stated that the optimum temperature to remove organic oil refinery pollutants was 45°C and the concentration of catalyst was 100 mg/L.

Combination of UV radiation and Fe(VI)

In this section of study, the combination of UV radiation and Fe(VI) oxidation process was investigated. The results showed that the increase in temperature up to 40°C had a positive effect and more than 40°C had a negative effect on dye removal efficiency using combination of UV radiation and Fe(VI) oxidation process (Figure 10b). The reason of this phenomenon was previously explained in the section of dye removal using Fe(VI) oxidation process alone. Also, the effect of pH on the dye removal efficiency showed that the highest dye removal efficiency was obtained at two pHs of 2 and 13 (Figure 10a). The results illustrated that the HRT between 0 and 10 minutes has a significant effect on the removal efficiency, but by increasing the HRT to more than 10 minutes, this effect became less significant. At 40°C, pH 2, Fe(VI) concentration of 3.3 mg/L, and the UV radiation power of 170 mW/cm², nearly 99% of the dye was removed in 15 min. However, the dye removal efficiency was not higher than 53% under the same condition at pH 9. Apparently, further studies are recommended to identify the reasons for such a behavior in the removal of dye using the combination of UV radiation and Fe(VI) oxidation process. Talaiekhosani et al reported that 65% and 73% of hydrogen sulfide and chemical oxygen demand (COD) were removed, respectively, when the combination of UV radiation and Fe(VI) was used (39). It seems that this method is more effective in the removal of dye compared with that of hydrogen sulfide and COD.

Combination of ZnO nanoparticles and Fe(VI) oxidation process

In this part of the study, the effect of pH and HRT

on the removal of dye using the combination of ZnO nanoparticles and the Fe(VI) oxidation process was investigated. The effect of pH between 2 and 13 is shown in Figure 11a. The pH reduction results in an increase in the removal efficiency of the dye. The maximum dye removal efficiency at pH 13 and HRT of 50 minutes was 28%. The maximum dye removal efficiency at HRT of 50 minutes was 57% when pH 9 was achieved. At pH of 2 and HRT of 50 minutes, the maximum removal efficiency was 67%. The optimum HRT was also different at various pH levels. For example, at pH values of 13, 9, and 2, the optimum HRT value was 20, 10, and 5 minutes, respectively.

The results of evaluation of the temperature effect on the dye removal efficiency using the combination of ZnO nanoparticles and the Fe(VI) oxidation process are shown in Figure 9b. The temperature increase had a positive effect on the dye removal efficiency using combination of ZnO nanoparticles and the Fe(VI) oxidation process. Increasing the temperature from 30 to 60°C led to an increase in the dye removal efficiency from 69% to nearly 100%. Honorio et al showed that 96% of the Reactive Blue 203 dye was removed from water using advance oxidation process (AOP) (42). Comparing these results with the results of this study revealed that combination of ZnO nanoparticles and Fe(VI) oxidation process may work better than the AOP.

Conclusion

In this study, the effective factors on the removal efficiency of Reactive Blue 203 using UV radiation, Fe(VI) oxidation process, absorption by ZnO nanoparticles, the combination of ZnO nanoparticles and Fe(VI) oxidation process, UV radiation and ZnO nanoparticles, and the composition UV radiation and Fe(VI) oxidation process were investigated. Also, the removal mechanism for some of the above-mentioned methods was presented. Removal of Reactive Blue 203 using Fe(VI) oxidation process can be completed during a fraction of seconds; therefore, it can be categorized as a rapid reaction. UV radiation, Fe(VI) oxidation process, and ZnO nanoparticles were able to remove 97%, 71%, and 47% of the dye under optimal conditions, respectively. Also, 100% of the dye was removed when the combination of Fe(VI) oxidation process and UV radiation was used under optimum conditions. Combination of ZnO nanoparticles and Fe(VI) oxidation process was also able to remove nearly 100% of the dye. Using the combination of ZnO nanoparticles and UV radiation under optimal conditions, the removal efficiency of the dye was close to 100%. All three factors including temperature, pH, and HRT were highly effective on hybrid processes. It was also found that by combining the various methods as mentioned in this study, the effect of some factors such as pH and temperature varies on the removal efficiency of dye using the processes. For example, when the ZnO nanoparticles alone were used, the increase in temperature have a negative effect on

the dye removal efficiency, but when the combination of UV radiation and ZnO nanoparticles was used, the temperature had a positive effect on the dye removal efficiency. Since the effect of the concentration of ZnO nanoparticles on the combination of ZnO nanoparticles and UV radiation was not investigated, it is recommended to investigate this parameter in future studies. The effect of dye concentration was not also evaluated in this study, therefore, it is recommended to investigate this parameter in future studies. Furthermore, it is recommended that the kinetics reaction of dye removal using UV radiation be evaluated in future studies.

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Ethical issues

Authors are aware of, and comply with, best practice in publication ethics specifically with regard to authorship (avoidance of guest authorship), dual submission, manipulation of figures, competing interests and compliance with policies on research ethics. Authors certify that submitted work is original and has not been published elsewhere.

Competing interests

The authors declare that there is no conflict of interests.

Authors' contribution

All authors were equally involved in the collection, analysis, and interpretation of the data. All authors critically reviewed, refined, and approved the manuscript.

References

- Hunge YM, Mahadik MA, Patil VL, Pawar AR, Gadakh SR, Moholkar AV, et al. Visible light assisted photoelectrocatalytic degradation of sugarcane factory wastewater by sprayed CZTS thin films. *J Phys Chem Solids* 2017; 111: 176-81. doi: 10.1016/j.jpics.2017.07.023.
- Talaiekhosani A, Banisharif F, Eskandari Z, Talaiekhosani MR, Park J, Rezaia S. Kinetic investigation of 1,9-dimethyl-methylene blue zinc chloride double salt removal from wastewater using ferrate (VI) and ultraviolet radiation. *Journal of King Saud University - Science* 2018. [In Press]. doi: 10.1016/j.jksus.2018.04.010.
- Yang B, Zuo J, Tang X, Liu F, Yu X, Tang X, et al. Effective ultrasound electrochemical degradation of methylene blue wastewater using a nanocoated electrode. *Ultrason Sonochem* 2014; 21(4): 1310-7. doi: 10.1016/j.ultrasonch.2014.01.008.
- Liu M, Chen Q, Lu K, Huang W, Lu Z, Zhou C, et al. High efficient removal of dyes from aqueous solution through nanofiltration using diethanolamine-modified polyamide thin-film composite membrane. *Sep Purif Technol* 2017; 173: 135-43. doi: 10.1016/j.seppur.2016.09.023.
- Jorfi S, Darvishi Cheshmeh Soltani R, Ahmadi M, Khataee A, Safari M. Sono-assisted adsorption of a textile dye on milk vetch-derived charcoal supported by silica nanopowder. *J Environ Manage* 2017; 187: 111-21. doi: 10.1016/j.jenvman.2016.11.042.
- Pirkarami A, Olya ME. Removal of dye from industrial wastewater with an emphasis on improving economic efficiency and degradation mechanism. *J Saudi Chem Soc* 2017; 21: S179-86. doi: 10.1016/j.jscs.2013.12.008.
- Ashtekar VS, Bhandari VM, Shirsath SR, Sai Chandra PL, Jolhe PD, Ghodke S. Dye wastewater treatment: Removal of reactive dyes using inorganic and organic coagulants. *Journal of Industrial Pollution Control* 2014; 30(1): 33-42.
- Isarain-Chavez E, Baro MD, Rossinyol E, Morales-Ortiz U, Sort J, Brillas E, et al. Comparative electrochemical oxidation of methyl orange azo dye using Ti/Ir-Pb, Ti/Ir-Sn, Ti/Ru-Pb, Ti/Pt-Pd and Ti/RuO₂ anodes. *Electrochim Acta* 2017; 244: 199-208. doi: 10.1016/j.electacta.2017.05.101.
- Tchobanoglous G, Burton FL, Stensel HD. *Wastewater engineering: treatment and reuse*. 4th ed. New York: McGraw-Hill; 2003.
- Eskandari Z. Control of hydrogen sulfide and organic compounds in municipal wastewater by using ferrate (VI) produced by electrochemical method [dissertation]. Isfahan: Jami Institute of Technology; 2016.
- Sharma VK. Ferrate(VI) and ferrate(V) oxidation of organic compounds: kinetics and mechanism. *Coord Chem Rev* 2013; 257(2): 495-510. doi: 10.1016/j.ccr.2012.04.014.
- Talaiekhosani A, Talaei MR, Rezaia S. An overview on production and application of ferrate (VI) for chemical oxidation, coagulation and disinfection of water and wastewater. *J Environ Chem Eng* 2017; 5(2): 1828-42. doi: 10.1016/j.jece.2017.03.025.
- Zhou Z, Fang S, Chen H, Ji J, Wu J. Trials of treating decentralized domestic sewage from a residential area by potassium ferrate(VI). *Water Air Soil Pollut* 2017; 228(8): 316. doi: 10.1007/s11270-017-3457-7.
- Sun S, Pang SY, Jiang J, Ma J, Huang Z, Zhang J, et al. The combination of ferrate(VI) and sulfite as a novel advanced oxidation process for enhanced degradation of organic contaminants. *Chem Eng J* 2018; 333: 11-9. doi: 10.1016/j.cej.2017.09.082.
- Sharma VK, Liu F, Tolan S, Sohn M, Kim H, Oturan MA. Oxidation of beta-lactam antibiotics by ferrate(VI). *Chem Eng J* 2013; 221: 446-51. doi: 10.1016/j.cej.2013.02.024.
- Rush JD, Bielski BH. Decay of ferrate (V) in neutral and acidic solutions. A premix-pulse radiolysis study. *Inorg Chem* 1994; 33(24): 5499-502. doi: 10.1021/ic00102a024.
- Rai PK, Lee J, Kailasa SK, Kwon EE, Tsang YF, Ok YS, et al. A critical review of ferrate(VI)-based remediation of soil and groundwater. *Environ Res* 2018; 160: 420-48. doi: 10.1016/j.envres.2017.10.016.
- Pepino Minetti RC, Macano HR, Britch J, Allende MC. In situ chemical oxidation of BTEX and MTBE by ferrate: pH dependence and stability. *J Hazard Mater* 2017; 324(Pt B): 448-56. doi: 10.1016/j.jhazmat.2016.11.010.
- Manoli K, Nakhla G, Ray AK, Sharma VK. Oxidation of caffeine by acid-activated ferrate(VI): effect of ions and natural organic matter. *AIChE J* 2017; 63(11): 4998-5006. doi: 10.1002/aic.15878.

20. Monfort O, Usman M, Soutrel I, Hanna K. Ferrate(VI) based chemical oxidation for the remediation of aged PCB contaminated soil: comparison with conventional oxidants and study of limiting factors. *Chemical Engineering Journal* 2019; 355(1): 109-17. doi: 10.1016/j.cej.2018.08.116.
21. Malik SN, Ghosh PC, Vaidya AN, Waindeskar V, Das S, Mudliar SN. Comparison of coagulation, ozone and ferrate treatment processes for color, COD and toxicity removal from complex textile wastewater. *Water Sci Technol* 2017; 76(5-6): 1001-10. doi: 10.2166/wst.2017.062.
22. Jorfi S, Barzegar G, Ahmadi M, Darvishi Cheshmeh Soltani R, Alah Jafarzadeh Haghighifard N, Takdastan A, et al. Enhanced coagulation-photocatalytic treatment of acid red 73 dye and real textile wastewater using UVA/synthesized MgO nanoparticles. *J Environ Manage* 2016; 177: 111-8. doi: 10.1016/j.jenvman.2016.04.005.
23. Rogers JA, Lagally MG, Nuzzo RG. Synthesis, assembly and applications of semiconductor nanomembranes. *Nature* 2011; 477(7362): 45-53. doi: 10.1038/nature10381.
24. Saravanan R, Sacari E, Gracia F, Khan MM, Mosquera E, Gupta VK. Conducting PANI stimulated ZnO system for visible light photocatalytic degradation of coloured dyes. *J Mol Liq* 2016; 221: 1029-33. doi: 10.1016/j.molliq.2016.06.074.
25. Goesmann H, Feldmann C. Nanoparticulate functional materials. *Angew Chem Int Ed Engl* 2010; 49(8): 1362-95. doi: 10.1002/anie.200903053.
26. Tian J, Zhao Z, Kumar A, Boughton RI, Liu H. Recent progress in design, synthesis, and applications of one-dimensional TiO₂ nanostructured surface heterostructures: a review. *Chem Soc Rev* 2014; 43(20): 6920-37. doi: 10.1039/c4cs00180j.
27. Djuricic AB, Chen X, Leung YH, Ng AM. ZnO nanostructures: growth, properties and applications. *J Mater Chem* 2012; 22(14): 6526-35. doi: 10.1039/c2jm15548f.
28. Jorfi S, Samaei MR, Darvishi Cheshmeh Soltani R, Talaie Khosani A, Ahmadi M, Barzegar G, et al. Enhancement of the bioremediation of pyrene-contaminated soils using a hematite nanoparticle-based modified fenton oxidation in a sequenced approach. *Soil Sediment Contam* 2017; 26(2): 141-56. doi: 10.1080/15320383.2017.1255875.
29. Darvishi Cheshmeh Soltani R, Jorfi S, Safari M, Rajaei MS. Enhanced sonocatalysis of textile wastewater using bentonite-supported ZnO nanoparticles: response surface methodological approach. *J Environ Manage* 2016; 179: 47-57. doi: 10.1016/j.jenvman.2016.05.001.
30. Darvishi Cheshmeh Soltani R, Rezaee A, Khataee AR, Safari M. Photocatalytic process by immobilized carbon black/ZnO nanocomposite for dye removal from aqueous medium: optimization by response surface methodology. *J Ind Eng Chem* 2014; 20(4): 1861-8. doi: 10.1016/j.jiec.2013.09.003.
31. Tan C, Gao N, Deng Y, Zhang Y, Sui M, Deng J, et al. Degradation of antipyrine by UV, UV/H₂O₂ and UV/PS. *J Hazard Mater* 2013; 260: 1008-16. doi: 10.1016/j.jhazmat.2013.06.060.
32. Talaiekhosani A, Eskandari Z, Talaie MR, Salari M. Hydrogen sulfide and organic compounds removal in municipal wastewater using ferrate (VI) and ultraviolet radiation. *Environ Health Eng Manag* 2017; 4(1): 7-14. doi: 10.15171/ehem.2017.02.
33. Vinayagam M, Ramachandran S, Ramya V, Sivasamy A. Photocatalytic degradation of orange G dye using ZnO/biomass activated carbon nanocomposite. *J Environ Chem Eng* 2018; 6(3): 3726-34. doi: 10.1016/j.jece.2017.06.005.
34. Zuorro A, Lavecchia R. Evaluation of UV/H₂O₂ advanced oxidation process (AOP) for the degradation of diazo dye reactive green 19 in aqueous solution. *Desalin Water Treat* 2014; 52(7-9): 1571-7. doi: 10.1080/19443994.2013.787553.
35. Zita J, Krysa J, Mills A. Correlation of oxidative and reductive dye bleaching on TiO₂ photocatalyst films. *J Photochem Photobiol A Chem* 2009; 203(2-3): 119-24. doi: 10.1016/j.jphotochem.2008.12.029.
36. Yang Y, Wyatt DT, Bahorsky M. Decolorization of dyes using UV/H₂O₂ photochemical oxidation. *Textile Chemist and Colorist* 1998; 30(4): 27-35.
37. Tichonovas M, Krugly E, Jankunaite D, Racys V, Martuzevicius D. Ozone-UV-catalysis based advanced oxidation process for wastewater treatment. *Environ Sci Pollut Res Int* 2017; 24(21): 17584-97. doi: 10.1007/s11356-017-9381-y.
38. Talaiekhosani A, Salari M, Talaie MR, Bagheri M, Eskandari Z. Formaldehyde removal from wastewater and air by using UV, ferrate(VI) and UV/ferrate(VI). *J Environ Manage* 2016; 184(Pt 2): 204-9. doi: 10.1016/j.jenvman.2016.09.084.
39. Talaiekhosani A, Eskandari Z, Bagheri M, Talaie MR. Removal of H₂S and COD using UV, ferrate and UV/ferrate from municipal wastewater. *Journal of Human, Environment and Health Promotion* 2016; 2(1): 1-8. doi: 10.29252/jhehp.2.1.1.
40. Kalyani DC, Telke AA, Dhanve RS, Jadhav JP. Ecofriendly biodegradation and detoxification of reactive red 2 textile dye by newly isolated *Pseudomonas* sp. SUK1. *J Hazard Mater* 2009; 163(2-3): 735-42. doi: 10.1016/j.jhazmat.2008.07.020.
41. Eskandari Z, Talaiekhosani A, Makipoor G, Jafari S, Rezaei S. Estimation of pollutants emission rate from activity of Isfahan city taxis. *J Air Pollut Health* 2017; 2(3): 137-44.
42. Honorio JF, Veit MT, Goncalves Gda C, de Campos EA, Fagundes Klen MR. Adsorption of reactive blue BF-5G dye by soybean hulls: kinetics, equilibrium and influencing factors. *Water Sci Technol* 2016; 73(5): 1166-74. doi: 10.2166/wst.2015.589.
43. Moussavi G, Mahmoudi M. Removal of azo and anthraquinone reactive dyes from industrial wastewaters using MgO nanoparticles. *J Hazard Mater* 2009; 168(2-3): 806-12. doi: 10.1016/j.jhazmat.2009.02.097.
44. Han Q, Wang H, Dong W, Liu T, Yin Y, Fan H. Degradation of bisphenol A by ferrate(VI) oxidation: kinetics, products and toxicity assessment. *Chem Eng J* 2015; 262: 34-40. doi: 10.1016/j.cej.2014.09.071.
45. Zhou Z, Jiang JQ. Reaction kinetics and oxidation products formation in the degradation of ciprofloxacin and ibuprofen by ferrate(VI). *Chemosphere* 2015; 119 Suppl: S95-100. doi: 10.1016/j.chemosphere.2014.04.006.
46. Li ZW, Gao W, Reeves RJ. Zinc oxide films by thermal oxidation of zinc thin films. *Surf Coat Technol* 2005; 198(1-3): 319-23. doi: 10.1016/j.surfcoat.2004.10.111.
47. Wang J, Gao L. Hydrothermal synthesis and photoluminescence properties of ZnO nanowires. *Solid*

- State Commun 2004; 132(3-4): 269-71. doi: 10.1016/j.ssc.2004.07.052.
48. Baratto C, Todros S, Faglia G, Comini E, Sberveglieri G, Lettieri S, et al. Luminescence response of ZnO nanowires to gas adsorption. *Sens Actuators B Chem* 2009; 140(2): 461-6. doi: 10.1016/j.snb.2009.05.018.
49. Ren S, Bai YF, Chen J, Deng SZ, Xu NS, Wu QB, et al. Catalyst-free synthesis of ZnO nanowire arrays on zinc substrate by low temperature thermal oxidation. *Mater Lett* 2007; 61(3): 666-70. doi: 10.1016/j.matlet.2006.05.031.
50. Shahrezaei F, Mansouri Y, Zinatizadeh AA, Akhbari A. Photocatalytic degradation of aniline using TiO₂ nanoparticles in a vertical circulating photocatalytic reactor. *Int J Photoenergy* 2012; 2012: 430638. doi: 10.1155/2012/430638.

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