

Application of Chitin and Zeolite adsorbents for treatment of low level radioactive liquid wastes

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Abstract

Two types of Shrimp Chitin derivatives and two types of Iranian natural Zeolite derivatives (Firuzkooch Clinoptilolite) were studied for adsorption and treatment of low-level radioactive liquid waste (LLW). Chitin with lower than 10% and Chitosan with higher than 90% deacetylation factor were selected as natural organic adsorbents. Natural Clinoptilolite of Firuzkooch area and Na form derivatives of it were selected as natural inorganic adsorbents. The static and dynamic ion exchange experimental results show that the adsorption efficiency depends on particle size, pH, adsorbent type, deacetylation factor (in Chitin adsorbents) and cation type. The best Cs adsorption occurred in Na form Clinoptilolite. Nevertheless Chitin derivatives, particularly Chitosan, are more efficient than Zeolite adsorbents for removing of radionuclides such as ¹³⁷Cs, ⁵⁴Mn, ⁹⁰Sr and ⁶⁰Co. Adsorption performance was discussed and compared with each other.

Key words: Chitin, Chitosan, Clinoptilolite, adsorption, radionuclides

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Introduction

Various kinds of adsorbents have been widely produced and applied for the removal of radionuclides and heavy metals (Kuribayashi et al., 1987). Among these, synthetic resins show high selectivity for the removal of metallic ions in liquid waste with high electric conductivity. Although this group of resins has excellent performance, the cost is relatively high in comparison with the ordinary type of ion exchange resins. This is due to the complexity of the synthetic producing process (Kuribayashi et al., 1988). Chitin has been widely used for cosmetics, food additives and so on due to it being a harmless and inexpensive material. Muzzarelli et al. showed the adsorption performance of Chitin. It has been shown that some kinds of Chitosan derivatives remove heavy metals from solutions (Muzzarelli, 1973, 1985, Muzzarelli et al., 1986). Zeolite is a naturally occurring mineral group consisting of over 50 different minerals (Zamzow et al., 1987). Made of a special crystalline structure that is porous but remains rigid in the presence of water, Zeolite can be adapted for a variety of use such as household odor control products, water and waste water treatment and so on (Kesraoul-Ouki, et al., 1990 and Cheeseman, et al., 1993 and Leppert, 1990).

It could be possible for the chitin and Zeolite derivatives to be applied to radioactive waste management if they show good performance for radionuclides removal in radioactive liquid waste (Cheeseman, et al., 1993 and Bailey, et al., 1999). Our objective is to evaluate the applicability of shrimp Chitin adsorbents compared to Iranian natural Clinoptilolite (from Firuzkooch area) to low-level radioactive liquid waste treatment.

Materials and Methods

Two types of Chitin and two types of Zeolite adsorbents were prepared for the static and dynamic experiments. Glass tube of inner diameter 2.5 cm filled with chitin adsorbents and 2 cm filled with Zeolite adsorbents were used as columns for dynamics experiments. Chitin and Chitosan were extracted from Shrimp shell waste with <10% and >90% deacetylation factor respectively (Ravi Kumar, 1999). Zeolite mines of Firuzkooch area prepared Iranian natural Clinoptilolite (Zeol1) and sodium form Clinoptilolite (ZeolNa). General characteristics of adsorbent are shown in Table 1. Three types of simulation solution were prepared to study and compare the adsorption performance. The chemical

composition is shown in Table 2. Solution group one is mono cationic and consists of stable isotopes. Solution group two consists of 3 stable cations of Cs, Sr and Co and solution group 3 is a simulated radioactive waste consisting of Cs-137, Co-60 and Mn-54. This solution was prepared by adding of Cs-137, Co-60 and Mn-54 to de-ionized water.

Gamma spectrometry was applied for a quantitative analysis of these radionuclides and ICP-AA used for measuring of stable radioisotopes.

In static experiments the solution volume was 25 ml. Weight of added adsorbent was 500 mg. The solution was put into glass beaker and mixed with the weighed adsorbent by magnetic stirrer. After definite period of time (Table 2) the samples were taken (2ml), it was centrifuged to be separated from the adsorbent and then 1 ml of it was placed in tubes for ICP-AA spectrophotometer. Distribution coefficient (K_d) was measured for each stable isotope. The K_d was calculated by equation (1):

$$K_{d (ml/g)} = [(A_o - A_{eq}) / A_{eq}] \times V/M \quad (1)$$

A_o : initial concentration of cation in solution

A_{eq} : concentration of cation after adsorption

V : solution volume (ml)

M : Weight of the adsorbent (g)

For obtaining the breakthrough curve the volume of the adsorbent bed was about 6 cm³. The volume of simulated radioactive waste was about 600 cm³. The input flow rate was about 1ml/min.

Results

Static methods

Figure 1 has shown maximum Distribution coefficient (K_d) of Chitin, Chitosan, Zeo1 and ZeoNa is treated by solution 1. Maximum adsorption of chitin adsorbents occurred within 45 minutes. Nevertheless major portion of these values belongs to the first 20 minutes and by increasing of treatment time over 45 minutes no effect on K_d was observed.

Chitin and Chitosan have shown maximum adsorption in Co solution and minimum in Sr solution. Increase in solution concentration, shifts this manner to Co>Sr>Cs. Optimum pH for Chitin adsorbent was 5 and lower or higher

pH has shown sensible decrease in K_d . The best adsorbent performance in Chitosan was seen in pH=6.5. However the best adsorption was seen at lower pH but some practical problems occurred due to solubility of Chitosan at pH lower than 6.5.

Zeolite adsorbents reached maximum adsorption in the longer treating time about 12-15 hours, with the major portion of adsorption occurred in 6 hours. Zeolite adsorption depended on pH solution and the optimum pH was 8.7. As shown in Figure 1 and 2, maximum adsorption belongs to Cs on Na form Clinoptilolite and minimum to Sr on natural Clinoptilolite. Clinoptilolite adsorption follows Cs>Co>Sr manner in both cases. By increasing concentration, K_d has been decreased but there was no observed effect on the adsorption manner. Figure 2 shows maximum adsorbent K_d with the treatment of solution 2. The treating time and optimal pH were similar to solution 1 cases. In this case, maximum K_d occurred nearly the same time as solution 1 and similarly decreasing of K_d occurred by the increasing of cation concentration. In Chitin adsorbents, cation adsorption by solution 2 followed the Co>Sr>Cs mode and in Clinoptilolite case, followed the same mode as solution 1.

Dynamic method

Figure 3 shows the breakthrough curve by chitin. ⁶⁰Co eluted first in about 10 bed volumes. In about 60 bed volumes it reaches to the breakthrough. ⁵⁴Mn starts to appear in 5 bed volumes, and by 70 bed volumes it reaches the breakthrough. In about 10 bed volumes ¹³⁷Cs starts to appear and by about 60 bed volumes it reaches to breakthrough points. The breakthrough curve for Chitosan is shown in Figure 4. As it could be expected from the K_d , Co was the best adsorbate and Cs, Mn as follows. Cs elution starts in 20 bed volumes by the column and at 100 bed volumes reached to 95% of breakthrough. The ⁶⁰Co started to appear in 40 bed volumes and 82% of breakthrough was completed with the 100 bed volumes. The Mn-54 started to appear in effluent at 30 bed volumes and reached to 94% at 100 bed volumes. The breakthrough curve for ZeoNa is shown in Figure 5. Cs was the best adsorbate in comparison with Co and Mn. Cs elution started in 20 bed volumes and at 100 bed volumes reached to 87% of breakthrough. The ⁶⁰Co and ⁵⁴Mn started to appear at 5 bed volumes. Both Mn, Co reach to full breakthrough about 40

Table 1: Characteristics of adsorbents

Chitin adsorbents characteristics	Zeolite adsorbents characteristics
<ul style="list-style-type: none"> - Fine fiber structure (20-50mesh or 297-840 microns) - Large specific surface - Ability to exchange cations - Low production cost - High resistively to chemical and radiation effects - Burnable without producing toxic substances 	<ul style="list-style-type: none"> - Fine particle with 50-100mesh (149-297 microns) - 3-dimensional rigid crystalline structure - Molecular sieve - Large specific surface - Ability to exchange cations - Low production cost - High resistively to radiation effects and low resistively to acidity condition

Table 2 : Treatment condition based on solution type

Properties Type of solution	Isotope type	pH range				Treatment Time		Experiment type
		Chitin	Chitosan	Zeo1	ZeoNa	Chitin Adsorbents (Min)	Zeo Adsorbents (Hr)	
Solution 1	Cs absolute solution	3-10	3-10	4.5-10.5	4.5-10.5	5,15,45	2,6,15	Batch
	Co absolute solution	3-10	3-10	4.5-10.5	4.5-10.5	5,15,45	2,6,15	Batch
	Sr absolute solution	3-10	3-10	4.5-10.5	4.5-10.5	5,15,45	2,6,15	Batch
Solution 2	Cs, Co, Sr mixture solution	5.5	6.7	8.7	8.7	5,15,30,60,120	2,6,12,24,36	Batch
Solution 3	¹³⁷ Cs, ⁶⁰ Co and ⁵⁴ Mn mixture solution	5.5	6.7	8.7	8.7	Each 60 min	Each one Hr	Column

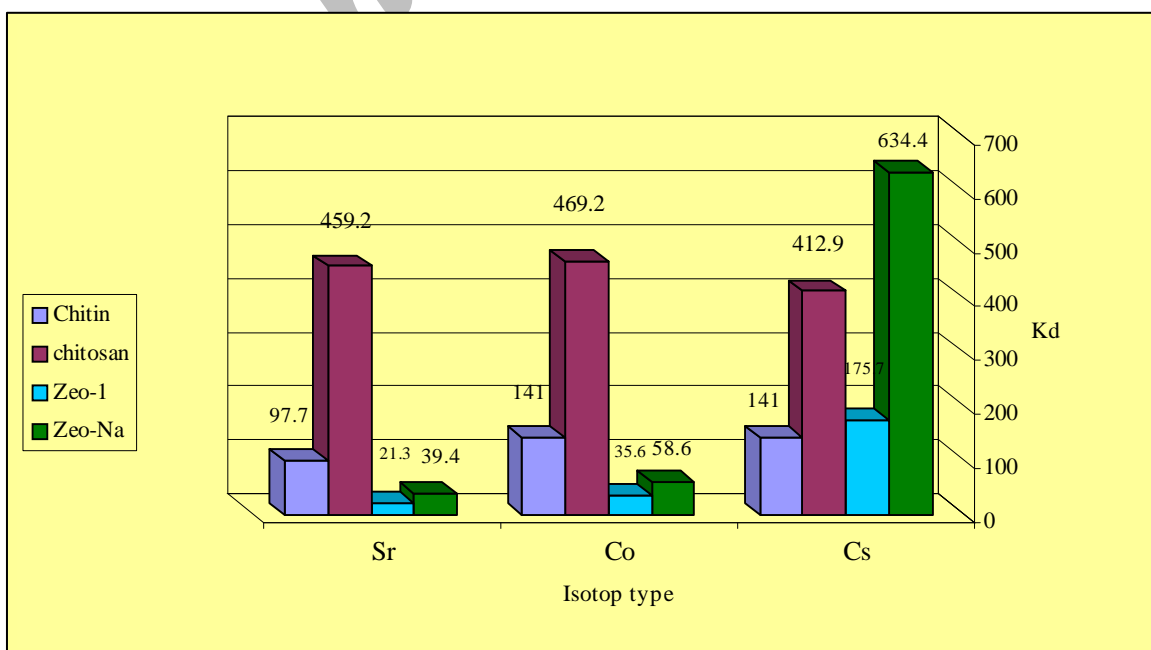


Figure 1: - Kd comparison chart based on adsorbent types and isotopes in solution 1

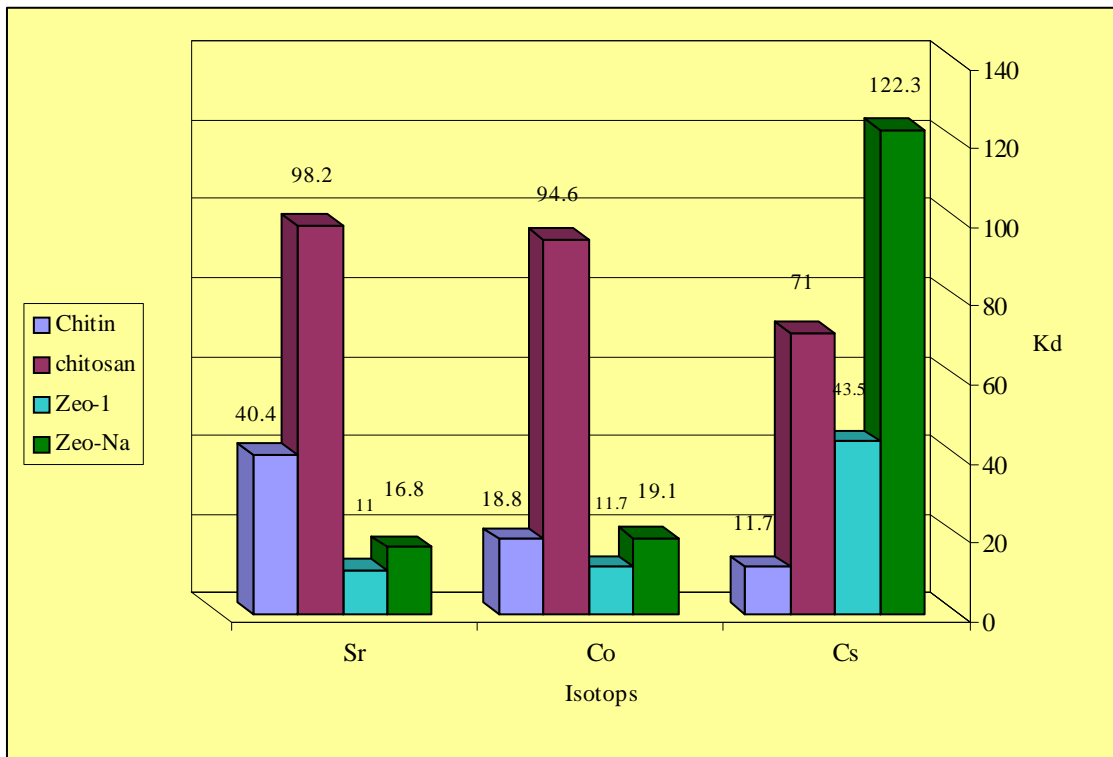


Figure 2: - Kd comparison chart based on adsorbent type and isotopes in solution 2

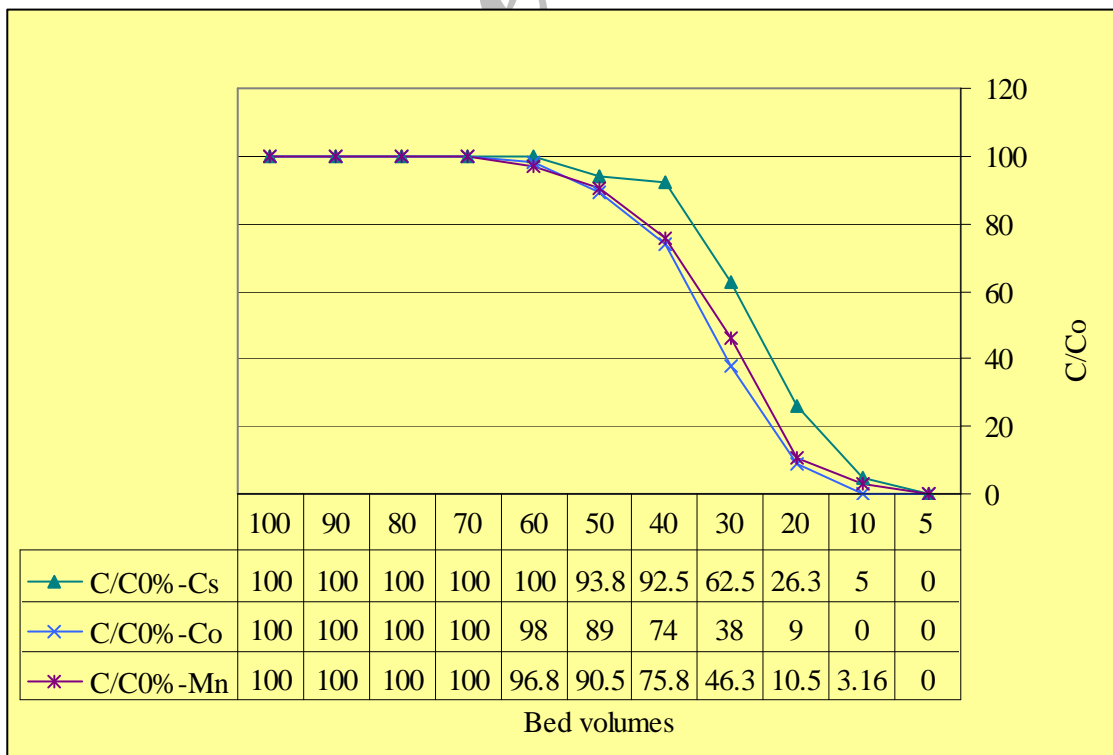


Figure 3: Break-through curve of Chitin

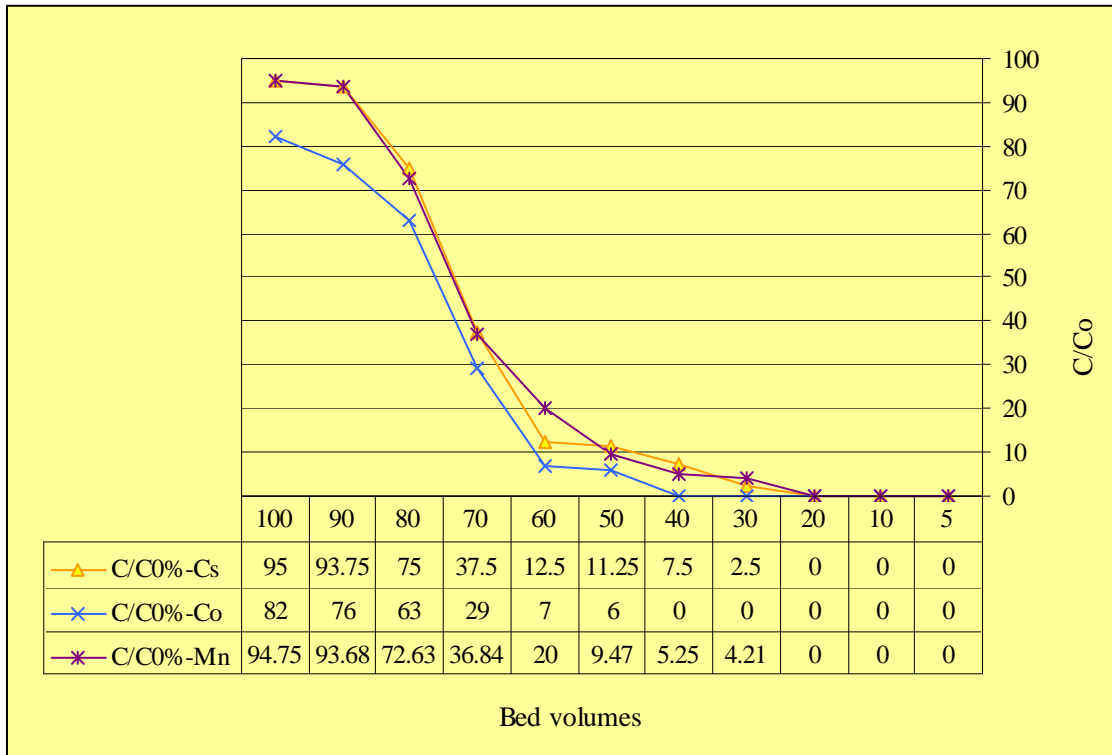


Figure 4: Break-through curve of Chitosan

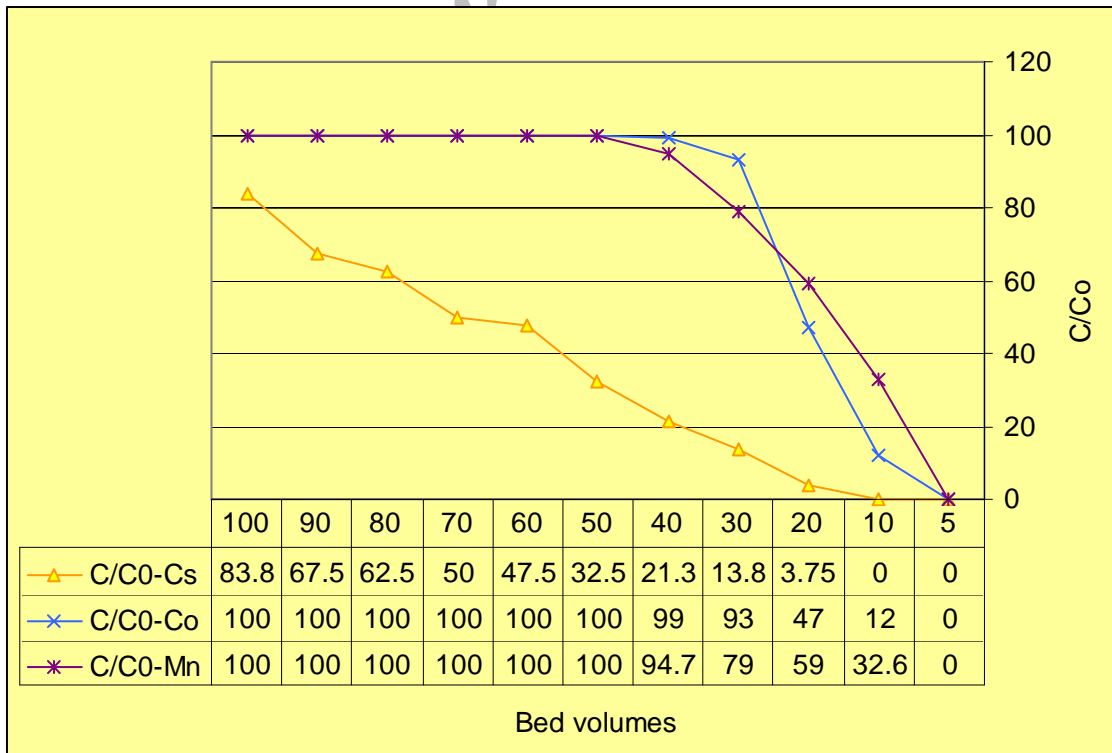


Figure 5: Break-through curve of ZeoNa

bed volumes. More or less natural Clinoptilolite showed no good effect on solutions and the experiments showed inconsistent results.

Discussion and Conclusion

Chitin and Chitosan adsorbents are effective for the removal of radionuclide. They can be used as suitable adsorbents for removing of radionuclide and heavy metals. Chitosan adsorbent had the best performance between four examined adsorbents. In this study natural Iranian Clinoptilolite (Firuzkooch area) showed the weakest performance among the adsorbents. Na form Clinoptilolite showed the best adsorption for Cs, and the less on Sr, Co, Mn ions.

In conclusion, Chitosan adsorbent can be applied for the treatment of radioactive liquid waste and water if suitable adsorbents were selected based on the solubility and deacetylation factor. Na form Clinoptilolite is suitable for removal of Cs from the solution, which solely contains it. Natural Clinoptilolite was not suitable for radionuclides removal.

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